Effects of Straw Incorporation on Soil Organic Matter and Soil Water-Stable Aggregates Content in Semiarid Regions of Northwest China

Peng Zhang^{1,2}⁹, Ting Wei^{1,2}⁹, Zhikuan Jia^{1,2}*, Qingfang Han^{1,2}, Xiaolong Ren^{1,2}, Yongping Li³

1 The Chinese Institute of Water-Saving Agriculture, Northwest A&F University, Yangling, Shaanxi, China, 2 Key Laboratory of Crop Physi-Ecology and Tillage Science in Northwestern Loess Plateau, Ministry of Agriculture, Northwest A&F University, Yangling, Shaanxi, China, 3 Guyuan Institute of Agricultural Sciences, Guyuan, Ningxia, China

Abstract

The soil degradation caused by conventional tillage in rain-fed areas of northwest China is known to reduce the water–use efficiency and crop yield because of reduced soil porosity and the decreased availability of soil water and nutrients. Thus, we investigated the effects of straw incorporation on soil aggregates with different straw incorporation rates in semiarid areas of southern Ningxia for a three-year period (2008–2010). Four treatments were tested: (i) no straw incorporation (CK); (ii) incorporation of maize straw at a low rate of 4 500 kg ha⁻¹ (L); (iii) incorporation of maize straw at a medium rate of 9000 kg ha⁻¹ (M); (iv) incorporation of maize straw at a high rate of 13 500 kg ha⁻¹ (H). The results in the final year of treatments (2010) showed that the mean soil organic carbon storage of the 0–60 cm soil layers were significantly (P<0.05) increased with H, M, and L, by 21.40%, 20.38% and 8.21% compared with CK, respectively. Straw incorporation increased >0.25 mm water-stable macroaggregates level, geometric mean diameter, mean weight diameter and the aggregate stability, which were ranked in order of increasing straw incorporation rates: H/M > L > CK. Straw incorporation significantly (P<0.05) reduced the fractal dimension in the 0–40 cm soil layers compared with CK. Our results suggest that straw incorporation is an effective practice for improving the soil aggregate structure and stability.

Citation: Zhang P, Wei T, Jia Z, Han Q, Ren X, et al. (2014) Effects of Straw Incorporation on Soil Organic Matter and Soil Water-Stable Aggregates Content in Semiarid Regions of Northwest China. PLoS ONE 9(3): e92839. doi:10.1371/journal.pone.0092839

Editor: Raffaella Balestrini, Institute for Plant Protection (IPP), CNR, Italy

Received July 19, 2013; Accepted February 26, 2014; Published March 24, 2014

Copyright: © 2014 Zhang et al. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Funding: This work was supported by the China Support Program (2012BAD09B03, 2011AA100504, and 2011BAD29B09) for Dryland Farming in the 12th 5-year plan period, the Youth project of National Natural Science Fund (31201156), and the Basic Science Research Fund in Northwest A&F University (QN2013005). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

* E-mail: jiazhk@126.com

• These authors contributed equally to this work.

Introduction

Soil infertility [1], soil erosion, and water deficiency [2] are the major factors that limited crop growth in semiarid areas of northwest China. The rates of crop straw use for fuel and forage have declined significantly since the 1980s and crop straw is increasingly burned after the harvest, which leads to high losses of soil organic substances [3,4], and increased emission of CO_2 that pollute the environment [5]. Furthermore, this practice has led to the degradation of the agricultural ecological environment [6].

The soil organic matter (SOM) content is one of the major factors that affects soil properties and functions including a range of physical characteristics such as the water-holding capacity [7], water infiltration [8], and aggregate stability [9]. SOM is considered to be a major binding agent that stabilizes soil aggregates [10,11]. Soil aggregates are the basic units of the soil structure [12], which are composed of primary particles and binding agents that determine the microbial biomass and mineral nutrient reserves [13–15]. These soil properties are also affected by soil organic matter decomposition processes [16,17].

Many studies have shown that crop straw is rich in organic material and soil nutrients, so it is increasingly considered to be an important natural organic fertilizer [18–20]. Straw can be

incorporated to soil either directly or indirectly, which can promote the production of a favorable soil environment. Straw also maintains the physicochemical condition of the soil and improves the overall ecological balance of the crop production system [20,21]. Nelson [22] and Wilhelm *et al.* [23] showed that the incorporation of crop residues into soil significantly prevented soil erosion and enhanced the soil quality. Sonnleitner *et al.* [24] found that straw incorporation also improved the aggregate stability and other soil properties compared with farmyard manure. Mulumba and Lal [25] also reported that the addition of crop residues to cultivated soil had positive effects on the soil porosity, available water content, soil aggregation, and bulk density. Bhagat and Verma [26] showed that the incorporation of crop straws for five years significantly increased the crop yield and improved the soil properties.

The soil improvement effect of straw incorporation has been recognized widely [23–25] but information is still limited on the responses of the SOC and water-stable aggregates under different rates of straw incorporation, particularly in the loessal soil in semiarid areas of northwest China. The theory and technique of straw incorporation in this region have also not been reported. Thus, the present study investigated the effects of different crop straw application rates combined with conventional planting on SOC, the >0.25 mm water-stable macroaggregate rate, and various soil properties in the southern Ningxia region of China.

Materials and Methods

Ethics Statements

The study was carried out on the private land, we rent the farmland from the local farmers, and contracts and deeds are signed. No specific permissions were required in this area to run the experiment as the study sites are farming area without any protection zone, and the farming activities won't hurt the local animals. And we only plant the grain crop in the field, so the field studies did not involve endangered or protected species.

Site description

The experiment was conducted between 2008 and 2010 at the Dryland Agricultural Research Station, Pengyang County, Ningxia, China (106°45'N, 35°79'E and 1800 m a.s.l.). The experimental area was in a hilly and gully region of the Loess Plateau, which was characterized by a semiarid, warm temperate, continental monsoon climate. The average annual rainfall was 435 mm, which fell mainly from June to September. The annual mean evaporation was 1050 mm and the annual temperature average was 8.1°C with a frost-free period of 155 days.

Rainfall during the experimental period was measured using an automatic weather station (WS-STD1, England) at the experimental site. Monthly precipitation distributions during the experimental period are shown in Fig. 1. The total precipitation for 2008, 2009, and 2010 was 390.9, 335.2, and 537 mm, while the precipitation during the maize-growing season was 362, 298.2, and 476.1 mm, respectively.

The soil at the experimental site was a loessal soil with a pH of 8.5. In the 0–40 cm soil layer, the organic matter, total N, P, and K were 8.32 g kg⁻¹, 0.61 g kg⁻¹, 0.58 g kg⁻¹, and 5.4 g kg⁻¹, respectively, while the available N, P, and K were 46.25 mg kg⁻¹, 10.41 mg kg⁻¹, and 104.82 mg kg⁻¹. In 2007, the site was planted with maize prior to the experiment

The experimental field was flat and, according to the FAO/ UNESCO Soil Classification [27], the soil was a Calcic Cambisol (sand 14%, silt 26%, and clay 60%) with low fertility. The key physical properties of the soil layers (0–40 cm depth) are shown in Table 1.

Experimental design and field management

The experiment used a randomized block design with three replicates. Each plot was 3 m wide and 6 m long. The experiment included four straw incorporation rate treatments: (i) no straw incorporation (CK); (ii) incorporation of maize straw at a low rate of 4 500 kg ha⁻¹ (L); (iii) incorporation of maize straw at a medium rate of 9000 kg ha⁻¹ (M); (iv) incorporation of maize straw at a high rate of 13 500 kg ha⁻¹ (H).

The maize straws were mixed manually with the top 25 cm of soil in the field. Before mixing with the soil, the maize straws were chopped into 5 cm pieces and then applied to the soil six months before the crop was planted to facilitate decomposition of the straw. The straw was incorporated into the soil layer on 15 October 2007 and after the crop harvests during 2008–2010.

Ten days before sowing, a basis fertilizer containing 102 kg N ha⁻¹ and 90 kg P ha⁻¹, was spread evenly over the each plot and plowed into soil layer. Maize (cv. Shendan 16) was sown at a rate of 5.25 seed m⁻² on 18 April 2008, 15 April 2009 and 20 April 2010 using a holesowing (3 cm in diameter) machine. An additional 102 kg N ha⁻¹ was applied as a top dressing in late June. And on 7 October 2008, 5 October 2008, and 10 October 2010. No irrigation was provided during the experimental years. Manual weeding was performed throughout the experiment.

Sampling and measurement

Rainfall data were recorded using a standard weather station located at the experimental site. After the maize harvest in 2008 and 2010, soil samples were collected for the four incorporation treatments. A soil sample was collected from each plot at depths of 0-20 cm, 20-40 cm and 40-60 cm to determine the soil organic matter. A similar soil sample was collected at depths of 0-10 cm, 10-20 cm, 20-30 cm and 30-40 cm to determine the aggregate stability. The soil samples were collected from four points in each plot replicate and mixed to produce a composite sample. Each soil samples was passed through an 8 mm sieve by gently breaking the soil clods, whereas pebbles and stable clods >8 mm were discarded. Soil samples were air-dried for 24 h in the laboratory before analysis.



Figure 1. Distribution of mean monthly precipitation at the experimental site during 2007–2010. doi:10.1371/journal.pone.0092839.g001

e 1. Physical properties of the tilth soil (0–40 cm depth) in the experimental site.
--

Depth (cm)	Bulk density (g cm ⁻³)	Aggregate size (%)					
		>5 mm	5–2 mm	2–1 mm	1–0.5 mm	0.5–0.25 mm	<0.25 mm
0–10	1.33	0.1	0.27	2.25	4.5	4.47	88.41
10–20	1.33	0.15	0.19	1.38	4.2	4.15	89.93
20–30	1.36	0.1	0.28	1.41	4.02	3.36	90.83
30–40	1.38	0.1	0.22	1.07	3.29	3.8	91.52

doi:10.1371/journal.pone.0092839.t001

Soil organic carbon was determined by the $K_2Cr_2O_7-H_2SO_4$ digestion method, and SOM content was calculated as a portion of SOC which has been described by Wang *et al.* [28].

$$OM_i = O_i \times A \tag{1}$$

$$C_i = \frac{d_i \times \rho_i \times O_i}{10} \tag{2}$$

where OM_i is the soil organic matter content (g kg⁻¹), O_i is the soil organic carbon content (g kg⁻¹), C_i is the soil organic carbon storage (Mg C ha⁻¹), d_i is the soil depth (cm), ρ_i is the soil bulk density (g cm⁻³), A is Van Bemmelen coefficient (A = 2) [29].

The size distribution of water-stable aggregates was determined by placing a soil sample on a stack of sieves (5,2, 1, 0.5 and 0.25 mm) fitted with a soil aggregate analyzer (Japan, QD24– DIK–2001). The stacked sieves were immersed in water and moved up and down by 3.5 cm at a frequency of 30 cycles $60s^{-1}$ for 15 min. The proportions of aggregates that measured >5, 5–2, 2–0.5, 0.5–0.25 and <0.25 mm were calculated [30].



Figure 2. Soil organic carbon storage with different straw incorporation treatments (Mg C ha⁻¹). Note: CK, no straw incorporation; L, incorporation of straw at a low rate of 4 500 kg ha⁻¹ maize straw; M, incorporation of straw at a medium rate of 9 000 kg ha⁻¹ maize straw; H, incorporation of straw at a high rate of 13 500 kg ha⁻¹ maize straw; Sum, the sum value of the 0–60 cm soil layers; Bars with different lower case letters indicate significant differences at P<0.05. doi:10.1371/journal.pone.0092839.g002

Table 2. WR_{0.25} with different straw incorporation treatments (%).

Year	Treatment	0-40 AVG				
		0—10	10—20	20—30	30—40	
2008	н	13.37a ₁ ±0.33 ^{B2}	10.57a±1.28 ₃ ^A	7.72ab±0.27 ^B	7.33a±0.51 ^A	9.75a±0.31 ^B
	М	11.79b±1.05 ^A	9.28a±0.60 ^B	7.97a±0.65 ^A	$6.62ab{\pm}0.85^{B}$	8.91b±0.26 ^B
	L	10.10c±0.46 ^A	7.39b±0.58 ^A	6.53bc±1.08 ^A	6.07ab±0.81 ^A	7.52c±0.16 ^A
	СК	9.12c±0.70 ^A	6.93b±0.78 ^A	5.89c±0.57 ^A	5.68b±0.63 ^A	6.91d±0.44 ^A
2009	Н	14.64a±0.49 ^{AB}	11.46a±0.17 ^A	9.51a±0.26 ^{AB}	7.85a±1.00 ^A	10.86a±0.19 ^{AB}
	М	14.93a±0.89 ^A	11.22a±0.14 ^{AB}	9.34a±0.32 ^A	7.06ab±1.46 ^{AB}	10.64a±0.27 ^{AB}
	L	$10.39b{\pm}0.26^{A}$	9.48b±1.49 ^A	7.02b±1.18 ^A	$6.48ab{\pm}1.10^{\text{A}}$	8.34b±0.32 ^A
	СК	9.75b±0.26 ^A	7.18c±0.73 ^A	6.40b±1.00 ^A	5.65b±0.35 ^A	7.25c±0.10 ^A
2010	Н	15.81a±1.20 ^A	12.76a±0.44 ^A	11.50a±0.20 ^A	8.57a±0.05 ^A	12.16a±0.59 ^A
	М	16.72a±1.36 ^A	12.81a±0.05 ^A	11.30a±1.10 ^A	8.81a±0.44 ^A	12.41a±0.72 ^A
	L	$11.00b{\pm}0.78^{A}$	9.06b±0.81 ^A	8.03b±2.14 ^A	6.96b±0.81 ^A	8.76b±0.66 ^A
	СК	10.23b±0.46 ^A	7.79b±0.23 ^A	6.07b±0.46 ^A	5.71c±0.23 ^A	7.45c±0.25 ^A

Note: CK, no straw incorporation; L, incorporation of straw at a low rate of 4 500 kg ha⁻¹ maize straw; M, incorporation of straw at a medium rate of 9 000 kg ha⁻¹ maize straw; H, incorporation of straw at a high rate of 13 500 kg ha⁻¹ maize straw; AVG, the mean value of the 0–40 cm soil layers.

¹Values followed by the same lowercase letter in the same line are not significantly different according to Duncan's multiple range test (*P*<0.05) between the four straw incorporation treatments in the same year.

²Values followed by the same uppercase letter in the same line are not significantly different according to Duncan's multiple range test (*P*<0.05) between the different years of the same straw incorporation treatment.

³Means \pm standard deviations.

doi:10.1371/journal.pone.0092839.t002

And the proportions of aggregates were used to calculate the water-stable macroaggregates content with a diameter of >0.25 mm [31], mean weight diameter (MWD) [32,33], geometric mean diameter (GMD) [34], and the soil aggregate stability (WSAR) [35]. These parameters were calculated as follows:

$$macroaggregates(>0.25mm) = \sum_{i=1}^{n} W_i$$
(3)

$$MWD = \sum_{i=1}^{n} \overline{X_i} W_i \tag{4}$$

$$GMD = \exp\left(\frac{\sum_{i=1}^{n} W_i \log \overline{X_i}}{\sum_{i=1}^{n} W_i}\right)$$
(5)

$$WSAR = WSA/A \times 100\% \tag{6}$$

where *macroaggregates*(>0.25*mm*) is the volume of soil particles with a diameter of >0.25 mm, W_i is weight of the aggregates in that size range as a fraction of total dry weight of the sample analysed, and *n* is number of sieves, $\overline{X_i}$ is the mean diameter of aggregates over each sieve size, *WSA* is the mass of *macroaggregates*(>0.25*mm*), and *A* is the mass of the soil aggregates with a diameter of >0.25 mm.

Fractal dimension D were then obtained to describe the characteristics of soil aggregate size distribution. As suggested by Tyler and Wheatcraft [36] and Zhang *et al.*, [37], the volume of soil with particle diameter $> d_i(d_i > d_{i+1}, i = 1, 2, ...)$ is defined as:

$$V(\delta > d_i) = A \left[1 - \left(d_i / k \right)^{3-D} \right) \right], \tag{7}$$

where δ is yard measure, k and A are the constants representing size and shape, respectively, and D is the fractal dimension. For a given soil, d_i represents the average value of soil particles between d_i and d_{i+1} .Generally, variations of particle density ρ among different soil particles could be ignored. And hence ρ is a constant. Therefore, another expression (Eq. (8)) is derived from Eq. (7):

$$W(\delta > d_i) = V(\delta > d_i)\rho = \rho A \left[1 - \left(d_i / k \right)^{3-D} \right) \right],$$
(8)

where $W(\delta > d_i)$ is the cumulative mass of particles with sizes $\delta > d_i$, and W_0 is the total mass of any sizes of soil particles. The fractal equation, reflecting the relationship between the mass distributions of soil particles and average particle diameter, can be obtained as follows:

$$\frac{W(\delta > d_i)}{W_0} = 1 - \left(\frac{\overline{d_i}}{d_{\max}}\right)^{3-D} or \ \frac{W(\delta > d_i)}{W_0} = \left(\frac{\overline{d_i}}{d_{\max}}\right)^{3-D} \tag{9}$$

Then after regression analysis between $\log(\overline{d_i}/\overline{d_{\max}})$ and $\log((W(\delta < \overline{d_i}))/W_0)$, the fractal dimension *D* can be calculated.

Statistical analysis

Statistical analysis was carried out by using the SPSS 13.0 (Statistical Package for the Social Sciences) package. The effects of treatments on the measured parameters were evaluated using a one-way ANOVA. Duncan's new multiple range test was used to calculate the least significant difference (LSD) between means



Figure 3. MWD values with wet sieving under the different straw incorporation treatments. Note: MWD, weight mean diameter; CK, no straw incorporation; L, incorporation of straw at a low rate of 4 500 kg ha⁻¹ maize straw; M, incorporation of straw at a medium rate of 9 000 kg ha⁻¹ maize straw; H, incorporation of straw at a high rate of 13 500 kg ha⁻¹ maize straw; Average, the mean value of the 0–40 cm soil layers. Bars with different lower case letters indicate significant differences at P<0.05. doi:10.1371/journal.pone.0092839.g003

when F-values were significant. In all cases, differences were deemed to be significant if $P{<}0.05$.

Results

Soil organic carbon storage (SOC)

The effect of straw incorporation on SOC storage is shown in Fig. 2, where the soil organic carbon storage increased with the straw incorporation. The sum of SOC storage in 0–60 cm layers with the three incorporation treatments were higher than CK, i.e., 7.71% (P<0.05), 11.14% (P<0.05) and 1.70% in 2008, 15.15%

(P < 0.05), 24.00% (P < 0.05) and 6.86% in 2009, and 21.40% (P < 0.05), 20.38% (P < 0.05) and 8.21% (P < 0.05) in 2010, respectively. The SOC storage increased with the number of years of incorporation, i.e., the SOC storage (0-60 cm depth) in 2010 had increased by 6.19–12.48% compared with 2008, and decreased with the soil layer depth, i.e., by 3.75–25.68% in 2008, 11.85–21.70% in 2009 and 13.51–26.64% in 2010. The SOC storage of H and M was slightly higher than CK in 2008, although the difference was significant in 40–60 cm layer only. In 2009, compared with CK, H and M significantly increased the SOC storage by 11.01% and 21.74% (P < 0.05) in 0–20 cm layer,



Figure 4. GMD values with wet sieving under the different straw incorporation treatments. Note: GMD, geometric mean diameter; CK, no straw incorporation; L, incorporation of straw at a low rate of 4 500 kg ha⁻¹ maize straw; M, incorporation of straw at a medium rate of 9 000 kg ha⁻¹ maize straw; H, incorporation of straw at a high rate of 13 500 kg ha⁻¹ maize straw; Average, the mean value of the 0–40 cm soil layers. Bars with different lower case letters indicate significant differences at P<0.05. doi:10.1371/journal.pone.0092839.g004

14.94% and 18.81% (P<0.05) in 20–40 cm layer, and 20.40% and 32.47% (P<0.05) in 40–60 cm layer, respectively, and L was significantly increased by 7.89% (P<0.05) in 40–60 cm layer only. There was a significant difference between straw incorporation treatments and CK for each of the three soil layers in 2010, but there were no significant differences between H and M.

Water-stable macroaggregates (>0.25 mm)

The effects of straw incorporation on >0.25 mm macroaggregates are shown in Table 2. The mean >0.25 mm macroaggregates contents of the 0–40 cm layers of incorporation treatments were significantly (P<0.05) higher than CK during 2008–2010, i.e., 8.91–41.14%, 15.13–49.94%, and 17.64–66.58%, respectively. The >0.25 mm macroaggregates content also decreased with the soil layer depth, and increased with the number of years of incorporation, i.e., the H, M, L and CK in 2010 was 24.72% (P<0.05), 39.28% (P<0.05), 16.49% and 7.81% higher than that in 2008. Compared with CK, the >0.25 mm macroaggregates content of three straw incorporation treatments in the 0–10 cm layer was increased by 10.75–46.60% in 2008, 6.56–53.13% in 2009, and 7.53–63.44% in 2010, respectively, while the differences between L and CK were not significant (P>0.05) during 2008–2010. The

Table 3. Effects on the soil aggregate stability rate with different straw incorporation treatments (%).

Year	Treatments	Soil Depth (cm)	0-40 AVG			
		0–10	10-20	20-30	30-40	
2008	Н	21.86a1±0.89 ^{A2}	16.29a±0.41 ₃ ^A	10.83a±0.54 ^B	9.19a±0.65 ^A	14.60a±0.33 ^A
	Μ	19.62ab±1.09 ^A	15.31b±0.62 ^A	11.16a±0.82 ^A	8.47b±0.21 ^A	13.70b±0.62 ^A
	L	17.83bc±1.56 ^A	12.88c±0.24 ^A	9.70b±0.64 ^A	8.00bc±0.42 ^A	11.83c±0.18 ^A
	СК	16.45c±2.42 ^A	12.52c±0.58 ^A	9.39b±0.80 ^A	7.53c±0.37 ^A	11.79c±0.45 ^A
2009	Н	$20.32a \pm 0.27^{A}$	16.14a±0.54 ^A	13.15a±0.34 ^{AB}	9.74a±0.99 ^A	14.84a±0.23 ^A
	Μ	$20.02a \pm 0.46^{A}$	15.41a±0.50 ^A	12.61a±0.41 ^A	8.66ab±1.28 ^A	14.18b±0.22 ^A
	L	15.76b±0.49 ^A	14.45a±1.78 ^A	10.16b±1.36 ^A	8.28ab±1.01 ^A	12.16c±0.35 ^A
	СК	15.47b±0.30 ^A	11.81b±0.83 ^A	9.59b±1.10 ^A	7.50b±0.41 ^A	11.09d±0.05 ^A
2010	н	20.19a±1.93 ^A	16.48a±1.73 ^A	15.24a±0.12 ^A	10.45a±0.11 ^A	15.59a±0.26 ^A
	Μ	$20.87a \pm 0.68^{A}$	15.78a±1.32 ^A	13.99a±1.23 ^A	10.08a±0.43 ^A	15.18a±0.64 ^A
	L	15.72b±0.72 ^A	13.39a±1.81 ^A	12.53b±0.46 ^A	8.66b±0.77 ^A	12.58b±1.00 ^A
	СК	15.11b±0.12 ^A	12.79a±1.70 ^A	12.00b±0.76 ^A	7.52c±0.26 ^A	11.85b±1.32 ^A

Note: CK, no straw incorporation; L, incorporation of straw at a low rate of 4 500 kg ha⁻¹ maize straw; M, incorporation of straw at a medium rate of 9 000 kg ha⁻¹ maize straw; H, incorporation of straw at a high rate of 13 500 kg ha⁻¹ maize straw; AVG, the mean value of the 0–40 cm soil layers.

¹Values followed by the same lowercase letter in the same line are not significantly different according to Duncan's multiple range test (*P*<0.05) between the four straw incorporation treatments in the same year.

²Values followed by the same uppercase letter in the same line are not significantly different according to Duncan's multiple range test (P<0.05) between the different years of the same straw incorporation treatment.

³Means \pm standard deviations.

doi:10.1371/journal.pone.0092839.t003

>0.25 mm macroaggregates content showed the same trends in the 10–20 and 20–30 cm layers, and the H and M treatments had the significant difference (P<0.05) with CK. The H, M and L increased by 29.05% (P<0.05), 16.55% and 6.87% in 2008, 38.94% (P<0.05), 24.96% and 14.69% in 2009, and 50.09% (P<0.05), 54.29% (P<0.05), 21.89% (P<0.05) in 2010, respectively. A linear correlation was found between the >0.25 mm macroaggregates and soil organic carbon (R^2 >0.64, P<0.01).

Mean weight diameter (*MWD*) and Geometric mean diameter (*GMD*)

Fig. 3 & 4 show that the MWD and GMD values with the three incorporation treatments increased significantly throughout the three-year study. The average MWD and GMD values under the incorporation treatments were higher in the 0-40 cm layers than CK during 2008-2010, i.e., 1.46-3.65% and 0.39-1.54% in 2008, 1.09-2.90% and 0.62-1.55% in 2009, 1.77-11.35% and 0.77-3.83% in 2010, respectively. There was no significant difference (P>0.05) between L and CK during the study period. The MWD and GMD values decreased with the soil layer depth and increased with the number of years of incorporation, i.e., the H, M, L and CK in 2010 was 10.56% (P<0.05) and 3.04% (P<0.05), 7.47% (P<0.05) and 2.29% (P<0.05), 3.24% and 1.15%, and 2.92% and 0.77% higher than that in 2008. The MWD and GMD values of treatments in the 0-10 and 10-20 cm layers were ranked in the order: H > M > L > CK. The H and M levels were significantly higher than CK while H and M levels were similar during 2008-2010 (Fig. 3-4). The MWD and GMD exhibited the same trends in 20-30 cm layers and there were no significant (P > 0.05) differences among incorporation treatments during 2008-2009, the MWD values with the H were significantly (P<0.05) higher than the other three treatments during 2010, i.e., by 5.50%, 10.04%, and 10.43%, respectively. The GMD did not differ significantly among the four treatments, with the exceptions of H and L, and H and CK in 2010. The MWD and GMD values in the 30-40 cm layers under the incorporation treatments were increased by 1.87%, 3.75% (*P*<0.05), 3.37% (*P*<0.05) and 0.76%, 1.76% (*P*<0.05), 0.78% in 2008, 0.37%, 0.01%, 0.73% and 0.73%, 0.45%, 0.45% in 2009, 8.09% (*P*<0.05), 7.35% (*P*<0.05), 2.21% (*P*<0.05) and 2.32% (*P*<0.05), 1.93% (*P*<0.05), 0.77% (*P*<0.05) in 2010, compared with CK, respectively. There was no significant difference (*P*>0.05) between H and M during the study period.

Soil aggregate stability (WSAR)

The straw incorporation significantly affected the WSAR after three years (Table 3). The WSAR values in the 0-40 cm layers of incorporation treatments was significantly higher than CK during 2008-2010, i.e., 0.34-19.80%, 9.65-33.79%, and 6.16-35.08%, respectively. The WSAR also decreased with the soil layer depth, while it increased with the number of years of incorporation. The WSAR in the 0-10 cm layers increased significantly with the amount of straw and the WSAR was highest with H throughout 2008–2010, i.e., the WSAR significantly (P < 0.05) increased by 28.35%, 28.93% and 25.78% compared with L, respectively, and by 33.62%, 31.35% and 32.89% (P<0.05) compared with CK, respectively. Compared with L and CK, the WSAR values with M increased significantly by 20.08-24.67% and 26.87-29.78% during 2008–2010, respectively, but there were no significant (P>0.05) differences between H and M, and L and CK. With the four treatments, the WSAR values in the 10-20 and 20-30 cm layers were lower than the 0–10 cm layers, which were ranked in the order: H > M > L > CK. There were no significant differences (P > 0.05) among the four treatments in the 30–40 cm layers throughout the three years, and the WSAR maintained at 7.50-10.45%.

Fractal dimension (D)

Fig. 5 shows that the fractal dimensions with the three incorporation treatments decreased significantly (P<0.05) after three years of straw incorporation (the R^2 of fitting curve is:



Figure 5. Fractal dimensions (D) of soil aggregates with different straw incorporation treatments. Note: CK, no straw incorporation; L, incorporation of straw at a low rate of 4 500 kg ha⁻¹ maize straw; M, incorporation of straw at a medium rate of 9 000 kg ha⁻¹ maize straw; H, incorporation of straw at a high rate of 13 500 kg ha⁻¹ maize straw; Average, the mean value of the 0–60 cm soil layers; Bars with different lower case letters indicate significant differences at P < 0.05; $0.92 < R^2 < 0.99$. doi:10.1371/journal.pone.0092839.g005

0.92~0.99). The fractal dimensions (0–40 cm layer) with the incorporation treatments were lower than those of CK during 2008–2010, i.e., 0.16% (P<0.05), 0.11% (P<0.05) and 0.02% in 2008, 0.14% (P<0.05), 0.16% (P<0.05) and 0.06% (P<0.05) in 2009, and 0.24% (P<0.05), 0.26% (P<0.05) and 0.06% in 2010, respectively. The fractal dimensions also increased with the soil layer depth and decreased with the number years of incorporation, i.e., the H, M, L and CK in 2010 was 0.20% (P<0.05), 0.27% (P<0.05), 0.17% (P<0.05) and 0.12% higher than that in 2008. The fractal dimensions in the 0–10 cm layers with the four treatments were ranked in the order: H<M<L<CK in 2008. The

fractal dimensions with H and M were significantly (P<0.05) lower than CK, i.e., 0.33%, 0.17% and 0.06%, respectively. However, the ranking was M<H<L<CK during 2009–2010, i.e., 0.06%, 0.20% (P<0.05), 0.02% in 2009, 0.35% (P<0.05), 0.42% (P<0.05) and 0.04% in 2010, respectively. While the differences between L and CK were not significant (P>0.05) during 2008–2010. Compared with L and CK, the fractal dimensions in the 10–20 cm layers under H was significantly (P<0.05) reduced the fractal dimensions by 0.11% and 0.14% in 2008, and 0.10% and 0.24% in 2009, while M significantly (P<0.05) reduced by 0.09% and 0.12% in 2008, and 0.08% and 0.23% in 2009, respectively.



Figure 6. Correlations between D and >0.25 mm macroaggregates, MWD, GMD, WSAR of the soil aggregates. Note: MWD, weight mean diameter of soil aggregates; GMD, geometric mean diameter of soil aggregates; WSAR, soil aggregate stability rate; D, Fractal dimension of soil aggregates. 95% Cl, 95% confidence interval.** *P*<0.01. doi:10.1371/journal.pone.0092839.g006

The fractal dimension was lowest with M during 2010, i.e., it was decreased by 0.22% (*P*<0.05) compared with L, and by 0.27% (*P*<0.05) compared with CK. The differences among the four treatments decreased gradually in the 20–40 cm layer and the fractal dimension remained at 2.987~2.991.

Data fitting detected a linear correlation between the indexes of soil aggregates and the fractal dimension (Fig. 6).

Discussion

Results of this study demonstrated that application of crop straws had positive effects on the soil physico-chemical properties. On the other hand, there are Loess plateau regions at northwest China with intensive cultivation systems and poor soil management strategies. Therefore, crop residual management is very important for preserving natural ecosystems [38]. Another problem for this area is conventional tillage that reduces the soil water storage and destroys the aggregates and soil structure [39] which prompts deterioration of crop yield.

The results indicated that the SOC storage increased significantly in the 0–60 cm layers after three years of straw incorporation (Fig. 2). Rasmussen and Collins [40] reported that the soil organic matter content was strongly related to the amount of residues added and only weakly related to the type of residue applied. The H and M treatments were significantly different from the L throughout the three years and the soil organic matter level increased as the straw incorporated and decomposed [20], which effectively mitigated the loss of soil organic carbon from in the agroecosystem caused by intensive cropping [41]. The SOC storage decreased gradually in all the treatments with the soil layer depth because the degree of straw incorporation was lower in the deeper layers compared with the surface layers (0–20 cm) of the soil [42]. This was because the amount of straw incorporated in the topsoil was greater than that in the deeper layers [40,43].

Soil aggregates are the basic units of the soil structure and they are composed of primary particles and binding agents [12]. They are also necessary soil conditions for high crop yields [44]. Conventional tillage disturbs the soil and increase the effects of drying-rewetting and freezing-thawing, which increases the susceptibility of the macroaggregate (>0.25 mm) to disruption [45–47]. Pinheiro *et al.* [48] showed that soil exposure with tillage and the lack of residue inputs led to a decline in aggregation and organic carbon, both of which made the soil susceptible to erosion. Our study showed that the straw incorporation of straws determined significantly more and larger soil aggregates than CK, thus indicating an improvement of soil physical quality. This may have been attributable to the significant increase in the SOC storage (an average increase by 8.21-21.40% in 2010), the lower soil bulk density (an average decreased by 1.80–4.13% in 2010, data no shown) [49], and increased soil porosity (an average increased by 1.70–3.90% in 2010, data no shown) [50] after straw incorporation. It also stimulated the activity of soil microorganisms [51] and an abundance of polyose metabolites were produced during the straw decay process [52]. These soil physical and chemical conditions may have accelerated the incorporated SOM decomposing process and increased soil aggregation [24].

Soil aggregate structural stability is widely recognized as a key indicator of soil quality, which is closely related to a number of soil properties, processes, and functions, e.g., the quantity and composition of SOM [48], infiltration capacity [53], soil biotic activity [54] and the resistance to erosion [55,56]. Wei et al. [57] showed that the addition of crop residues was the most effective measure for increasing the rhizosphere aggregate stability. Sonnleitner et al. [24] and Karami et al. [41] also found that straw application improved the aggregate stability and other soil properties. In our study, the soil aggregate stability of the straw incorporation treatments were significantly higher than CK in 2010, and it was decreased with the soil layer depth. These results agreed with studies by Tripathy and Singh [42] and Karami et al. [41]. Our results also indicated that straw incorporation was positively related to the physical protection of organic matter [20] and an increased aggregate quantity [58], but it also improved the soil aggregate stability [20] and reduced soil degeneration [22,23].

Many studies have shown that soil is a porous medium with fractal characteristics [59–60]. Thus, fractal theory can be used to describe the complex characteristic of soil structure [61]. Castrignanò and Stelluti [62] reported that a higher fractal dimension indicated the heavier texture of a soil and its inferior permeation properties. This showed that fractal theory is an effective method to describing the soil aggregate distribution [37] and changed with different levels of straw incorporation [31]. The fractal dimensions of the 0–40 cm layers with the four treatments

References

- Hu KL, Li BG, Lin QM, Li GT, Chen DL (1999) Spatial variability of soil nutrient in wheat field. Transactions of the Chinese Society of Agricultural Engineering 15:33–38. (in Chinese with English abstract)
- Mupangwa W, Twomlow S, Walker S (2008) The influence of conservation tillage methods on soil water regimes in semi-arid southern Zimbabwe. Phys Chem Earth, Parts A/B/C 33: 762–767.
- Biederbeck V, Campbell C, Bowren K, Schnitzer M, McIver R (1980) Effect of burning cereal straw on soil properties and grain yields in Saskatchewan. Soil Sci Soc Am J 44: 103–111.
- Wuest SB, Caesar-TonThat T, Wright SF, Williams JD (2005) Organic matter addition, N, and residue burning effects on infiltration, biological, and physical properties of an intensively tilled silt-loam soil. Soil Tillage Res 84: 154–167.
- Duan F, Liu X, Yu T, Cachier H (2004) Identification and estimate of biomass burning contribution to the urban aerosol organic carbon concentrations in Beijing. Atmos Environ 38: 1275–1282.
- Mandal KG, Misra AK, Hati KM, Bandyopadhyay KK, Ghosh PK, et al. (2004) Rice residue-management options and effects on soil properties and crop productivity. J Food Agric Environ 2: 224–231.
- 7. Carter MR (2002) Soil quality for sustainable land management. Agron J 94: 38–47.
- Hillel D (2004) Encyclopedia of Soils in the Environment: London, Elsevier, v. 4, p. 295–303.
- Six J, Bossuyt H, Degryze S, Denef K (2004) A history of research on the link between (micro) aggregates, soil biota, and soil organic matter dynamics. Soil Tillage Res 79: 7–31.
- Tisdall J, Oades J (1982) Organic matter and water-stable aggregates in soils. J Soil Sci 33:141–163. doi:10.1111/j.1365-2389.1982.tb01755.x
- 11. Haynes R, Beare M (1997) Influence of six crop species on aggregate stability and some labile organic matter fractions. Soil Biol Biochem 29: 1647–1653.
- Scanlon BR, Andraski BJ, Bilskie J (2002) Methods of Soil Analysis: Part 4 Physical Methods: Soil Science Society of America. 643–668 p.
- Hernández-Hernández R, López-Hernández D (2002) Microbial biomass, mineral nitrogen and carbon content in savanna soil aggregates under conventional and no-tillage. Soil Biol Biochem 34: 1563–1570.

after three years were ranked in the order: $H \le M \le L \le CK$ and the three straw incorporation treatments were significantly different from CK. These results agreed with Zhang *et al.* [37] and Zhang *et al.* [31]. The low values of D indicated a size distribution dominated by a large number of macroaggregates (>0.25 mm) [59,63]. This improvement in the fractal dimension may have been accelerated by the incorporation of straw, which improved the soil structure, increased the SOM content and microbial activity [64], and significantly increased the mount and size of soil aggregates [37]. Our results indicated there were significant improvements in the soil macroaggregates and the aggregate structure after straw incorporation [20].

Conclusion

The incorporation of different amount of straw significant increased the SOC storage, >0.25 mm macroaggregates, MWD and GMD in a semiarid soil. The SOC storage, >0.25 mm macroaggregates, MWD and GMD also increased with higher straw incorporation rates. The fractal dimension decreased with increasing straw incorporation rates. Therefore, the incorporation of straw into the soil in semiarid areas is an effective practice for improving the soil aggregate content and stability.

Acknowledgments

We are grateful to Mr Li Yongping and Liu Shixin for managing the field experiments and professional English editor Jackson who is from UK, and kind help with the language of this manuscript.

Author Contributions

Conceived and designed the experiments: ZJ QH. Performed the experiments: PZ TW. Analyzed the data: PZ. Contributed reagents/ materials/analysis tools: XR YL. Wrote the paper: PZ.

- Villar M, Petrikova V, Dnaz-Ravina M, Carballas T (2004) Changes in soil microbial biomass and aggregate stability following burning and soil rehabilitation. Geoderma 122: 73–82.
- Ashagrie Y, Zech W, Guggenberger G (2005) Transformation of a Podocarpus falcatu dominated natural forest into a monoculture Eucalyptus globulus plantation at Munesa, Ethiopia: soil organic C, N and S dynamics in primary particle and aggregate-size fractions. Agric Ecosyst Environ 106: 89–98.
- Jastrow J (1996) Soil aggregate formation and the accrual of particulate and mineral-associated organic matter. Soil Biol Biochem 28: 665–676.
- Chevallier T, Blanchart E, Albrecht A, Feller C (2004) The physical protection of soil organic carbon in aggregates: a mechanism of carbon storage in a Vertisol under pasture and market gardening (Martinique, West Indies). Agric Ecosyst Environ 103: 375–387.
- Duiker S, Lal R (1999) Crop residue and tillage effects on carbon sequestration in a Luvisol in central Ohio. Soil Tillage Res 52: 73–81.
- Saroa G, Lal R (2003) Soil restorative effects of mulching on aggregation and carbon sequestration in a Miamian soil in central Ohio. Land Degrad Dev 14: 481–493.
- Tan DS, Jin JY, H SW, Li ST, He P (2007) Effect of long-term application of K fertilizer and wheat straw to soil on crop yield and soil K under different planting systems. Agric Sci in China 6: 200–207.
- Li Q, Chen X, Zhang F, Römheld V (2002) Study on balance of phosphorus and potassium in winter wheat and summer maize rotation system. Plant Nutr Fert Sci 8: 152–156. (in Chinese with English abstract)
- Nelson RG (2002) Resource assessment and removal analysis for corn stover and wheat straw in the Eastern and Midwestern United States—rainfall and windinduced soil erosion methodology. Biomass Bioenergy 22: 349–363.
- Wilhelm W, Johnson JMF, Karlen DL, Lightle DT (2007) Corn stover to sustain soil organic carbon further constrains biomass supply. Agron J 99: 1665–1667.
- Sonnleitner R, Lorbeer E, Schinner F (2003) Effects of straw, vegetable oil and whey on physical and microbiological properties of a chernozem. App Soil Ecol 22: 195–204.
- Mulumba LN, Lal R (2008) Mulching effects on selected soil physical properties. Soil Tillage Res 98: 106–111.
- Bhagat R, Verma T (1991) Impact of rice straw management on soil physical properties and wheat yield. Soil Sci 152: 108–115.

- FAO/UNESCO (1993) 'World soil resources. An explanatory note on the FAO World Soil Resource Map at 1 : 25,000,000 scales.' (FAO: Rome)
- Wang S, Tian H, Liu J, Pan S (2003) Pattern and change of soil organic carbon storage in China: 1960s–1980s. Tellus B 55: 416–427.
- Pribyl DW (2010) A critical review of the conventional SOC to SOM conversion factor. Geoderma 156: 75–83.
- Oades J, Waters A (1991) Aggregate hierarchy in soils. Soil Res 29: 815–828.
 Zhang P, Jia ZK, Wang W, Lu WT, Gao F, et al. (2012) Effects of straw returning on characteristics of soil aggregates in semi-arid areas in southern Ningxia of China. Sci Agric Sin 45:1513-1520. (in Chinese with English abstract)
- Van Bavel C (1949) Mean weight diameter of soil aggregates as a statistical index of aggregation. Soil Sci Soc Am Proc. pp. 20–23.
- Youker R, McGuinness J (1957) A short method of obtaining mean weightdiameter values of aggregate analyses of soils. Soil Sci 83: 291–294.
- Mazurak AP (1950) Effect of gaseous phase on water-stable synthetic aggregates. Soil Sci 69: 135–148.
- Hou XQ, Li R, Jia ZK, Han QF, Wang W, et al. (2012) Effects of rotational tillage practices on soil properties, winter wheat yields and water-use efficiency in semi-arid areas of north-west China. Field Crops Res 129: 7–13.
- Tyler SW, Wheatcraft SW (1992) Fractal scaling of soil particle-size distributions: analysis and limitations. Soil Sci Soc Am J 56: 362–369.
- Zhang Z, Wei C, Xie D, Gao M, Zeng X (2008) Effects of land use patterns on soil aggregate stability in Sichuan Basin, China. Particuology 6: 157–166.
- Zhu H, Wu J, Huang D, Zhu Q, Liu S, et al. (2010) Improving fertility and productivity of a highly-weathered upland soil in subtropical China by incorporating rice straw. Plant Soil 331: 427–437.
- Mele PM, Crowley DE (2008) Application of self-organizing maps for assessing soil biological quality. Agric Ecosyst Environ 126: 139–152.
- Rasmussen PE, Collins HP (1991) Long-term impacts of tillage, fertilizer, and crop residue on soil organic matter in temperate semiarid regions. Adv Agron 45: 93–134.
- Karami A, Homaee M, Afzalinia S, Ruhipour H, Basirat S (2012) Organic resource management: Impacts on soil aggregate stability and other soil physicochemical properties. Agric Ecosyst Environ 148: 22–28.
- Tripathy R, Singh AK (2004) Effect of water and nitrogen management on aggregate size and carbon enrichment of soil in rice-wheat cropping system. J Plant Nutr Soil Sci 167: 216–228.
- Prasad R, Power J (1991) Crop residue management. Adv Soil Sci 15: 205–251.
 Limon-Ortega A, Govaerts B, Sayre KD (2009) Crop rotation, wheat straw
- management, and chicken manure effects on soil quality. Agron J 101: 600–606.
 Beare M, Hendrix P, Cabrera M, Coleman D (1994) Aggregate-protected and unprotected organic matter pools in conventional-and no-tillage soils. Soil Sci Soc Am J 58: 787–795.
- Paustian K, Levine E, Post WM, Ryzhova IM (1997) The use of models to integrate information and understanding of soil C at the regional scale. Geoderma 79: 227–260.

- Mikha MM, Rice CW (2004) Tillage and manure effects on soil and aggregateassociated carbon and nitrogen. Soil Sci Soc Am J 68: 809–816.
- Pinheiro E, Pereira M, Anjos L (2004) Aggregate distribution and soil organic matter under different tillage systems for vegetable crops in a Red Latosol from Brazil. Soil Tillage Res 77: 79–84.
- Lal R (2000) Physical management of soils of the tropics: priorities for the 21st century. Soil Sci 165: 191–207.
- Pagliai M, Vignozzi N, Pellegrini S (2004) Soil structure and the effect of management practices. Soil Tillage Res 79: 131–143.
- Kasteel R, Garnier P, Vachier P, Coquet Y (2007) Dye tracer infiltration in the plough layer after straw incorporation. Geoderma 137: 360–369.
- Pascual JA, García C, Hernandez T (1999) Comparison of fresh and composted organic waste in their efficacy for the improvement of arid soil quality. Bioresour Technol 68: 255–264.
- Abu-Sharar T, Salameh A (1995) Reductions in hydraulic conductivity and infiltration rate in relation to aggregate stability and irrigation water turbidity. Agric Water Manage 29: 53–62.
- Roldán A, Salinas-García J, Alguacil M, Caravaca F (2005) Changes in soil enzyme activity, fertility, aggregation and C sequestration mediated by conservation tillage practices and water regime in a maize field. Appl Soil Ecol 30: 11–20.
- Barthes B, Roose E (2002) Aggregate stability as an indicator of soil susceptibility to runoff and erosion; validation at several levels. Catena 47: 133–149.
- Ramos M, Nacci S, Pla I (2003) Effect of raindrop impact and its relationship with aggregate stability to different disaggregation forces. Catena 53: 365–376.
- Wei CF, Gao M, Shao JG, Xie DT, Pan GX (2006) Soil aggregate and its response to land management practices. China Particuology 4: 211–219.
- Tarafdar JC, Meena SC, Kathju S (2001) Influence of straw size on activity and biomass of soil microorganisms during decomposition. Eur J Soil Biol 37: 157– 160.
- Perfect E, Kay B (1991) Fractal theory applied to soil aggregation. Soil Sci Soc Am J 55: 1552–1558.
- Rieu M, Sposito G (1991) Fractal fragmentation, soil porosity, and soil water properties: II. Applications. Soil Sci Soc Am J 55: 1239–1244.
- Bird N, Bartoli F, Dexter A (2005) Water retention models for fractal soil structures. Eur J Soil Sci 47: 1–6.
- Castrignanò A, Stelluti M (1999) Fractal geometry and geostatistics for describing the field variability of soil aggregation. J Agric Eng Res 73: 13–18.
- Ding Q, Ding W (2007) Comparing stress wavelets with fragment fractals for soil structure quantification. Soil Tillage Res 93: 316–323.
- Liu JF, Hong W, Wu CZ (2002) Fractal features of soil clusters under some precious hardwood stands in the central subtropical region, China. Acta Ecol. Sin. 22, 197. (in Chinese with English abstract).